$See \ discussions, stats, and author \ profiles \ for \ this \ publication \ at: \ https://www.researchgate.net/publication/373048321$

Historic sampling of a vanishing beast: population structure and diversity in the black rhinoceros

Article in Molecular Biology and Evolution · August 2023

DOI: 10.1093/molbev/msad180

CITATIONS	3	READS	
0		43	
16 autho	ors, including:		
	Binia De Cahsan		Michael V Westbury
	University of Copenhagen		GLOBE institute
	21 PUBLICATIONS 147 CITATIONS		89 PUBLICATIONS 817 CITATIONS
	SEE PROFILE		SEE PROFILE
	Xin Sun		Ashot Margaryan
	Peking University		University of Copenhagen
	12 PUBLICATIONS 143 CITATIONS		66 PUBLICATIONS 2,648 CITATIONS
	SEE PROFILE		SEE PROFILE

Some of the authors of this publication are also working on these related projects:



rice epigenomics View project

Project African Small Mammal Phylogenetics/Phylogeography View project

Historic sampling of a vanishing beast: population 2 structure and diversity in the black rhinoceros 3 4 Fátima Sánchez-Barreiro^{1,16}, Binia De Cahsan^{1,16,*}, Michael V. Westbury¹, Xin Sun¹, Ashot 5 Margaryan¹, Claudia Fontsere^{2,1}, Michael W. Bruford^{3,†}, Isa-Rita M. Russo³, Daniela C. Kalthoff⁴, 6 Thomas Sicheritz-Pontén^{1,5}, Bent Petersen^{1,5}, Love Dalén^{6,7}, Guojie Zhang⁸⁻¹¹, Tomás Marquès-7 Bonet^{2,12,13}, M. Thomas P. Gilbert^{1,14,17,*}, Yoshan Moodley^{15,17} 8 9 ¹Globe Institute, University of Copenhagen, Øster Voldgade 5-7, 1350 Copenhagen, Denmark 10 11 ²Institut de Biologia Evolutiva (Consejo Superior de Investigaciones Científicas-Universitat Pompeu Fabra), Barcelona 12 Biomedical Research Park, Doctor Aiguader 88, 08003 Barcelona, Catalonia, Spain. 13 ³Cardiff School of Biosciences, Sir Martin Evans Building, Cardiff University, Museum Avenue, Cardiff, 14 CF10 3AX, United Kingdom ⁴Department of Zoology, Swedish Museum of Natural History, Frescativägen 40, 114 18 Stockholm, Sweden 15 16 ⁵Centre of Excellence for Omics-Driven Computational Biodiscovery (COMBio), Faculty of Applied Sciences, AIMST 17 University, Kedah, Malaysia ⁶Centre for Palaeogenetics, Svante Arrhenius väg 20C, 10691 Stockholm, Sweden 18 19 ⁷Department of Bioinformatics and Genetics, Swedish Museum of Natural History, Frescativägen 40, 114 18 Stockholm, 20 Sweden 21 ⁸Section for Ecology and Evolution, Department of Biology, University of Copenhagen, 2100 Copenhagen, Denmark 22 ⁹State Key Laboratory of Genetic Resources and Evolution, Kunming Institute of Zoology, Chinese Academy of Sciences, 23 650223, Kunming, People's Republic of China 24 ¹⁰Center for Excellence in Animal Evolution and Genetics, Chinese Academy of Sciences, 650223, Kunming, People's 25 Republic of China 26 ¹¹BGI-Shenzhen, 518083, Shenzhen, People's Republic of China ¹²National Centre for Genomic Analysis-Centre for Genomic Regulation, Barcelona Institute of Science and Technology, 27 08028 Barcelona, Spain. 28 29 ¹³Institucio Catalana de Recerca i Estudis Avançats (ICREA), 08010 Barcelona, Catalonia, Spain. 30 ¹⁴NTNU University Museum, 7491 Trondheim, Norway 31 ¹⁵Department of Biological Sciences, University of Venda, University Road, 0950 Thohoyandou, Republic of South Africa 32 ¹⁶Co-first authors 33 ¹⁷Senior Authors 34 [†]Deceased 35 *Correspondence: binia.cahsan@sund.ku.dk, tgilbert@sund.ku.dk (M.T.P.G.) © The Author(s) 2023. Published by Oxford University Press on behalf of Society for Molecular Biology and Evolution.

1

© The Author(s) 2023. Published by Oxford University Press on behalf of Society for Molecular Biology and Evolution. This is an Open Access article distributed under the terms of the Creative Commons Attribution-NonCommercial License (https://creativecommons.org/licenses/by-nc/4.0/), which permits non-commercial re-use, distribution, and reproduction in any medium, provided the original work is properly cited. For commercial re-use, please contact journals.permissions@oup.com

1 Abstract

2 The black rhinoceros (Diceros bicomis L.) is a critically endangered species historically distributed across sub-3 Saharan Africa. Hunting and habitat disturbance have diminished both its numbers and distribution since the 19th 4 century, but a poaching crisis in the late 20th century drove them to the brink of extinction. Genetic and genomic 5 assessments can greatly increase our knowledge of the species and inform management strategies. However, when 6 a species has been severely reduced, with the extirpation and artificial admixture of several populations, it is 7 extremely challenging to obtain an accurate understanding of historic population structure and evolutionary history 8 from extant samples. Therefore, we generated and analysed whole-genomes from 63 black rhinoceros museum 9 specimens collected between 1775 and 1981. Results showed that the black rhinoceros could be genetically structured into six major historic populations (Central Africa, East Africa, Northwestern Africa, Northeastern 10 11 Africa, Ruvuma and Southern Africa) within which were nested four further subpopulations (Massailand, Southwestern, Eastern Rift and Northern Rift), largely mirroring geography, with a punctuated north-south cline. 12 However, we detected varying degrees of admixture among groups, and found that several geographical barriers, 13 most prominently the Zambezi River, drove population discontinuities. Genomic diversity was high in the middle 14 15 of the range and decayed toward the periphery. This comprehensive historic portrait also allowed us to ascertain the ancestry of 20 re-sequenced genomes from extant populations. Lastly, using insights gained from this unique 16 17 temporal dataset, we suggest management strategies, some of which require urgent implementation, for the conservation of the remaining black rhinoceros diversity. 18

19 Introduction

Next generation DNA sequencing technology is finding increasing application in conservation management (Shafer et al. 2015). Until recently however, the majority of population scale conservation genomic studies have utilised reduced representation methods, which call single nucleotide polymorphisms (SNPs) from a limited set of randomly amplified loci (Hohenlohe et al. 2021). By comparison, whole genome sequences allow for precise estimates of mutation and recombination rates; higher resolution insights into population diversity, structure, demography and evolutionary history; and with the benefit of positional information that allows the detection and timing of introgression, inbreeding and, as sample sizes increase, local adaptation (Theissinger et al. 2023).

These attributes of whole genome sequences make them an indispensable tool for managers entrusted with the conservation of the planet's remaining biodiversity. Conservation practice is reliant on all knowledge available for the biodiversity under protection, but as an important starting point, the species population structure, or at a minimum, a subspecies level taxonomy is essential (Coates et al. 2018). Genetic data are, in fact, crucial when defining management units, such as evolutionarily significant units (ESUs), and for estimating levels of genetic diversity, inbreeding and gene flow, all of which guide conservation decisions (Barbosa et al. 2018).

34

35 In the present study we applied whole genome resequencing across a temporally distributed dataset, with the aim

of filling existing knowledge gaps related to a critically endangered African megaherbivore, the black rhinoceros
(*Diceros bicornis* L.). Prior to 1960, the black rhinoceros had been the most abundant extant rhinoceros species,
although its population had started to decline in the 19th century due to habitat clearance and unsustainable hunting
(Emslie 2020). As of 2021, some 6,195 black rhinoceroses were left across the continent (Ferreira et al. 2022),
reflecting a modest, yet positive demographic recovery after the lowest recorded census size of 2,354 animals in
the early 1990's. This historic low was the result of a ca. 98% decline of the wild population between 1960 and
1995, owing principally to intense poaching for the rhinoceros horn trade (Emslie 2020).

8

9 The historic distribution of the black rhinoceros encompassed a vast, continuous area across sub-Saharan Africa, 10 that spanned a broad range of habitats, from bushland and grassland, to desert, only avoiding areas of dense tropical rainforest (Rookmaaker and Antoine 2012; Figure 1). Currently outside of zoos, the species survives almost solely 11 12 in a few protected areas, with large (> 1,000 individuals) managed metapopulations only in South Africa and Namibia. In Kenya, the black rhinoceros has made a steady recovery from no more than 381 individuals in 1987, 13 to 897 by 2021 (Kenya Wildlife Service 2021). In Zimbabwe it has made a slower recovery from ca. 300 14 individuals in 1995 (Kotzé et al. 2014) to 616 in 2021 (Ferreira et al. 2022). Apart from these four countries, 15 Tanzania is the only remaining country with aboriginal populations of black rhinoceros, however, the present-day 16 population of no more than 160 is poorly managed, scattered across a handful of reserves and stem from a 17 minimum estimate of 31 in 1995 (Emslie and Brooks 1999). Small satellite populations of black rhinoceros have 18 been reestablished in some former range states, but these comprise a total population of about 212 (Ferreira et al. 19 2022) and have used animals mostly from South Africa. Therefore, the persistence of the black rhinoceros in what 20 21 remains of its heavily fragmented natural range is heavily dependent on active conservation efforts, which include population genetic management (Moodley et al. 2017). 22

23

24 The subspecies level taxonomy of the black rhinoceros has been contentious among rhinoceros experts for over a century (Rookmaaker 2011). In the late 1980s, a pragmatic classification into four "ecotypes" was settled by the 25 26 African Elephant and Rhino specialist group (AERSG, du Toit 1987) to aid conservation efforts. These were the 27 southwestern black rhinoceros of Namibia, the south-central black rhinoceros ranging from South Africa to Tanzania, the eastern black rhinoceros of East Africa, and the now-extinct western black rhinoceros from West 28 29 Africa. Despite a lack of supporting taxonomic evidence and ignoring other more detailed assessments by Groves 30 (1967) and Zukowsky (1965), the AERSG classification has persisted until present times. This is problematic because the four ecotypes are managed separately and are often incorrectly and misleadingly referred to as 31 32 subspecies.

Genetic assessments are key sources of information for determining how populations are structured across the species distribution. In this regard, although a substantial body of prior work exists for the black rhinoceros, these have mostly focused on either single ecotypes, or a subset of the managed populations (Harley et al. 2005; Karsten et al. 2011; Muya et al. 2011; Van Coeverden de Groot et al. 2011; Anderson-Lederer et al. 2012; Kotzé et al. 2014). The species- and range-wide understanding of the population structure and diversity has been less well explored. Furthermore, given the major recent population extirpations and bottlenecks that the black rhinoceros has experienced, its current population structure and diversity may not be an accurate reflection of what existed
just half a century ago. Thus, an improved understanding of its pre-decline status will be essential for both
expanding our knowledge of its ecology and evolution, but also informing future conservation efforts. In this
regard, despite the extirpation of the species across much of its historic distribution, obtaining a representative
range-wide genetic sample is possible thanks to the wealth of historic specimens preserved in museum collections.

- 7 This temporal sampling approach to study black rhinoceros genetics was explored for the first time by Moodley 8 et al. (2017), who investigated the species-level population structure, phylogeny, and genetic erosion through time. 9 However, their analyses were limited to molecular data from mitochondrial DNA and microsatellite markers. 10 Although their data showed a major population genetic break on either side of the Zambezi River, most of their 11 conclusions about genetic structure and diversity outside southern Africa were based on the history of the 12 mitochondrial control region, since only a fraction of their historic samples yielded enough microsatellite data.
- 13

Therefore, in this study we aimed to expand the resolution and scope of this previous work by taking advantage 14 of palaeogenomic sequencing techniques. Specifically, we generated whole-genome re-sequencing data for a set 15 of historic black rhinoceros specimens, representing most of the species' historic distribution, and supplemented 16 this with genomic data from a number of individuals from extant populations. Ultimately, our goal was to use 17 these genomes to resolve the patterns of population structure, gene flow and diversity in the black rhinoceros prior 18 to their decline in order to better inform conservation management. In parallel, we aimed to evaluate whether 19 modern individuals are still representative of historic individuals from the same geographic region and could 20 therefore provide the basis for the recovery of historic populations by informing future range expansion efforts. 21

22 **Results**

23 A black rhinoceros whole-genome temporal dataset

We generated shotgun DNA sequencing data for 98 individual black rhinoceroses originally sampled from sixteen countries across the historic and contemporary range of the species. The historic specimens (n = 71) ranged in collection date between 1775 and 1981. The initial 27 modern samples derived from extant populations in natural reserves: one Namibian (Etosha National Park), three Kenyan (Maasai Mara Game Reserve, Nairobi National Park and Ol Pejeta Conservancy), and two South African (iMfolozi and Mkhuze Game Reserves; Figure 1). However, based on our relatedness analysis (Figure S1) we excluded seven individuals from our modern data set from further downstream analyses.

31

We mapped the raw sequence data against the publicly available whole-genome assembly for the black rhinoceros ASM1363453v1 (Genbank Assembly Accession: GCA_013634535.1; (Moodley et al. 2020). We excluded individuals with depth of coverage < 1x from further analyses (n=8), yielding a final dataset consisting of 63 historic and 20 modern unrelated, re-sequenced whole genomes. Importantly, 53 of the historic genomes

- 1 corresponded to samples whose associated metadata included coordinates indicating geographic origin (Figure 1).
- 2 3

Historic samples were named as follows: the alpha-2 code of the country of origin (see Table 1), the year of

4 collection, and an index number (to distinguish samples of identical country and year). Country of origin was

5 unknown for two samples, which we indicated by replacing the country code with 'un'. Modern samples were

- 6 labelled with simpler identifiers that included the country code (for South Africa and Namibia) or reserve code
- 7 (for Kenya) followed by an index number (see Table 1 for further details on the distribution of samples across
 8 countries).
- 9

As expected, the DNA sequence data from the historic specimens showed signals characteristic of ancient DNA, including cytosine deamination, shorter library insert sizes and sizable fractions of non-endogenous DNA (Figure S2). As such, the average depth of coverage for the nuclear genomes were generally lower and more variable among the historic specimens (ranging between 1.27x and 20.11x), compared to modern samples (ranging between 7.37x and 22.78x; Table S1). The endogenous DNA of the historic samples ranged from 5% to 62% (Table S1 and Figure S2), and levels of cytosine deamination ranged from 0.5% to 5%.

16

17 Table 1. Overview of the number and origin of black rhinoceros samples in the dataset. For historic and modem 18 samples separately, the countries of origin and their corresponding alpha-2 codes are specified, as well as the 19 number of re-sequenced genomes and the historic populations present in each country.

Country / reserve	Code	Re-sequenced genomes	Populations (K=6)	Subpopulations (K=10)
Angola	AO	5	S	SW, SN/SE
Botswana	BW	1	S	SN/SE
Chad	TD	3	NW	NW
DRC	CD	2	EA, RU	EA, RU
Ethiopia	ET	2	NE	NE
Kenya	KE	19	EA, CE	CE, EA, MA, ER, NR
Malawi	MW	2	RU	RU
Mozambique	MZ	1	CE	ER
Nigeria	NG	1	NW	NW
Somalia	SO	5	NE	NE
South Africa	ZA	2	S	SN/SE
South Sudan	SS	3	EA	EA
Tanzania	TZ	8	CE, RU	MA, ER, RU
Uganda	UG	1	EA	NR
unknown	un	2	CE, S	CE, SN/SE
Zambia	ZM	3	CE	CE
Zimbabwe	ZW	3	S	SN/SE
TOTAL		63		
Maasai Mara Game Reserve	МА	7	Modern CE-EA	MA

Nairobi National Park	NNP	3	Modern CE-EA	Modern CE-EA
Ol Pejeta Conservancy	OP	14	Modern CE-EA	Modern CE-EA
Etosha National Park, Namibia	NA	1	Modern S	SW
iMfolozi and Mkhuze, South Africa	ZA	2	Modern S	SN/SE
TOTAL		27		

1 Black rhinoceroses exhibited geography-driven population structure

We used a range-wide data set comprising 63 historic genomes to determine the population structure of the black
rhinoceros prior to its decline in the late 20th century. We used genotype likelihoods of variant transversions as

4 input for the following population structure analyses (see *Variant site identification* in Methods and Figure S3).

5

6 We first performed a principal component analysis (PCA) to explore historic population structure. The first 7 principal component separated southern African samples from the rest: individuals from south and west of the 8 Zambezi River, that is southern Angola, Namibia, Botswana, Zimbabwe, and South Africa, were clearly grouped 9 apart from central, eastern and northern samples (Figure 2A). The second principal component separated 10 individuals sampled in Chad and Nigeria from others in northeastern, eastern and Central Africa.

11

12 To investigate these patterns of structure in finer detail, we conducted PCA analyses of the individuals on either side of the Zambezi River separately. We observed five major populations north and east of the river (Figure 2B), 13 14 largely clustering according to geography: the northwestern population (Chad and Nigeria, NW) observed in 15 Figure 2A; a northeastern population (NE) from Ethiopia and Somalia; an East African population (EA) including animals from South Sudan, Uganda, north DRC (Democratic Republic of Congo) and north western Kenya; a 16 Central African population (CE) from southern Kenya, northern and central Tanzania, Zambia, and Mozambique; 17 18 and a more distinct population localised to Malawi, southeastern Tanzania and putatively from the southern DRC, previously suggested by mtDNA and named Ruvuma (RU, Moodley et al. 2017). 19

20

The samples from south of the Zambezi River also displayed substructure, but along an east-west axis. Individuals from South Africa (including our Cape rhinoceros sample), Zimbabwe, Botswana and south eastern Angola were separated (along PC1) from those originating in south western Angola (Figure 2D). PC2 separated South Africa and Zimbabwe from Botswana and southeastern Angola (Figure 2D).

25

We conducted analogous PCA analyses including the genomes of 20 unrelated modern samples (see *Relatedness test* in Methods and Figure S1) on either side of the Zambezi River separately. Modern samples from the Kenyan reserves fell within and between the historic EA and CE samples, while the Namibian modern individual grouped with the historic southwestern Angola samples, and the South African modern genome grouped among the historic Zimbabwe-South Africa individuals. Therefore, the observed subpopulation groupings within southern Africa follow closely the three subpopulations, SW (Namibia and southwestern Angola), SN (southeastern Angola,

- 1 Botswana, Zimbabwe), and SE (South Africa and Zimbabwe) previously identified by Moodley et al. (2017).
- 2

3 To investigate historic population structure in more detail, we used the 63 historic genomes in an admixture proportion analysis and observed a pattern largely concordant with the results of the PCA analyses (Figure 2F and 4 5 Figure S4). The value of K = 6 was found to be the most likely for the data set using EvalAdmix (Figure S5). At 6 K = 2, as with PC1 in Figure 2A, individuals from southern Africa (S) separated from those north and east of the 7 Zambezi River, although RU appeared to comprise a mixture of alleles from both populations. As K increased, 8 NW, NE and EA separated from CE and RU at K = 3, NW was then distinguished from NE and EA at K = 4. 9 Then, RU separated from CE at K = 5, while EA separated from NE at K = 6 (Figure 2F). Higher K models also 10 yielded similar EvalAdmix results as well as geographically interesting and conservation relevant subpopulation structure. At K = 7 (Figure S4), a subpopulation, closely related to CE, could be distinguished among five genomes 11 sampled in the Maasailand region in the rift valley of southern Kenya and northern Tanzania, previously identified 12 from mtDNA as Chari-Victoria (CV, Moodley et al. 2017). However, the higher resolution offered by whole 13 genomes placed the three individuals sampled on the Chari River into the NW population, making the name CV 14 in appropriate for the Massailand genomes, which we renamed here MA. At K = 8, SW was delineated from SN/SE. 15 K = 9 separated five further genomes from southern Kenya and Tanzania, but the range of this subpopulation did 16 not overlap with MA, instead these individuals were sampled in the relatively narrow gap to the east of the rift 17 valley and to the west of the distribution of RU. Thus, while MA is characteristic of black rhinoceros in Maasailand 18 and the southern rift valley, this new subpopulation is more associated with the area to the east of the rift valley, 19 and so we name it here ER. Finally, at K = 10, four genomes from a region including Uganda, Lake Turkana and 20 Lake Baringo were differentiated from EA. We name the subpopulation of black rhinoceros inhabiting this arid 21 landscape NR, as it is dominated by volcanoes and lakes of the northern rift valley. 22

23

The distinctive range-wide population structuring at K = 6 also allowed the detection of admixture (Figure 2F). Some individuals within EA and CE were not assigned fully to either population, instead appearing admixed. In Kenya and South Sudan two EA genomes showed admixture with NE and CE, and one EA genome from Uganda appeared admixed with NW, while another in DRC was admixed with CE. Three CE genomes from Kenya were admixed with EA, two of which were the only samples from the valley of the Tana River in our data set (see Table S1). Fourteen CE individuals were admixed with RU at K = 6, but at K = 7, the most admixed of these were designated MA.

31

Analogously to our PCA analyses, we also conducted an admixture analysis including the 20 modern unrelated individuals (Figures S6, S7). At K = 6, modern individuals from two Kenyan reserves, Ol Pejeta Conservancy and Nairobi National Park, appeared either fully EA in ancestry, or as admixed between EA and CE (Figure S6). On the other hand, our three samples from the Maasai Mara Game Reserve showed a high proportion of MA ancestry (K = 8, Figure S6), which is geographically consistent because the reserve is situated within the Maasailand region, with some EA that is absent in historic MA genomes. The Namibian sample showed ancestry from the SW subpopulation, while the South African modern individual clustered with the historic SN/SE subpopulation (Figure 1 S6).

2

3 Lastly, to determine the relationships among populations and timing of key divergence events, we reconstructed a fossil calibrated genome-wide phylogeny using one, least admixed, individual for each of the populations 4 5 identified at K = 6 (Figure 2G). We used a sliding window approach with 20 kb windows and a 1 Mb slide. As in 6 the above analyses, the most supported topology featured an initial split between southern Africa (S) and other 7 regions. The only exception was that RU was a sister lineage with S and not to other genomes from East Africa. 8 Within the Eastern clade, NW branched before CE, while NE and EA were the most derived sister lineages. 9 Although we recovered high bootstrap values (100) for all nodes, both the gene and site concordance factors were 10 low, with maximum values of 27.3 and 39.8 respectively, suggesting high levels of phylogenetic discordance in our dataset (Figure 2G). Using an estimate for the divergence of the black and white rhinoceros species from a 11 12 common ancestor of between 5.3 and 7.3 million years (Ma), we inferred the first population split to have occurred between 0.73 - 1.22 Mya, with all other major population subdivisions likely occurring before ~500 kya. 13

14

The observed levels of structuring at increasing values of K prompted a more explicit test of whether a model of 15 isolation by distance (IBD) might have driven the population structure of the black rhinoceros in historic times. 16 We therefore conducted a Mantel test (Mantel 1967) on our 53 georeferenced historic samples, which revealed a 17 significant correlation between genomic and geographical distance, with geography potentially explaining up to 18 68% of the total variation in the data set (Figure 3A). The pairwise distances obtained when comparing our one 19 20 individual from South Africa (ZA1775.1) to any other individuals in our dataset, were markedly higher relative to other comparisons. This individual was not only highly geographically isolated, but also temporally isolated, as it 21 was sampled from 1775, compared to 1845-1981 for other historic samples. However, this sample was also the 22 23 lowest coverage (1.27x) which could have also driven relatively higher levels of divergence. We suspected that 24 these higher pair-wise values could increase the significance of our Mantel regression, however, the test remained significant even when this individual was removed from the analysis (Figure S8). We further investigated the 25 26 effect of sampling date with genetic distance but found only a very weak correlation (Figure S9).

27

We also calculated D-statistics to determine whether EA and CE individuals, inhabiting the middle of the species 28 29 range in central and eastern Africa, were closer to NW or to S, located at the extremes of the range. Under an 30 isolation by distance scenario, the expectation would be a linear decline in D-statistic values with distance from the centre of the range. For this, we investigated shared derived polymorphisms using the topology (((S, NW), 31 32 EA|CE), Outgroup). The D-statistic is commonly used for assessments of gene flow, that is, assuming the input 33 topology corresponds to the correct phylogenetic tree (Figure 2G). However in this case, S and NW are not sister 34 populations, and so elevated D-scores in this analysis will reflect shared polymorphisms due to closer common 35 ancestry as opposed to gene flow (Westbury et al. 2018; Westbury et al. 2021). Thus, a negative D-score would 36 indicate a closer relationship of the test group (EA or CE) to the S population, whereas a positive D-score would 37 indicate a closer relationship to the NW population. We observed a decline in D-statistics as the distance to the 38 northwestern end of the range increased (Figure 3B). Interestingly, however, the decline was not as linear as

1 expected (Figure 3B). In fact, three groups of samples were identifiable through this approach, and they matched

2 the EA, CE and RU individuals as sorted by the PCA and admixture analyses (Figure 3B).

3

To further investigate these potentially varying degrees of connectivity among the historic black rhinoceros 4 5 populations, we explored range-wide gene flow (or barriers to it) by computing effective migration rates with 6 EEMS (Petkova et al. 2016) using the 53 georeferenced historic genomes. The resulting effective migration surface 7 pinpointed areas where genetic differentiation decayed quickly with distance (higher than average effective 8 migration, blue-shaded, Figure 4), versus areas where genetic differentiation remained high even in relatively close 9 geographic space (lower than average effective migration, grey-shaded, Figure 4). Importantly, low effective migration might be due to an actual barrier to gene flow, or to low population density in the area (Petkova et al. 10 2016). 11

12

We observed broad regions of low effective migration for the black rhinoceros across sub-Saharan Africa. These 13 included the central Congo basin, where the species never occurred, extending south approximately through the 14 valleys of the Kafue and Lower Zambezi to the Indian Ocean (I, Figure 4), but also up the Shire valley into the 15 basin of Lake Malawi and from there into the Kilombero and Rufiji valleys of south-central Tanzania (II). From 16 17 the Congo basin, this low effective migration surface also extended both east roughly through the basin of Lake 18 Victoria, across the Gregory Rift and along the Tana River valley to the Indian Ocean (III), and north into Central 19 African Republic and Sudan, along the Bahr-al-Ghazal and southeast through the valleys of the White and the Blue Nile, across the Ethiopian Rift and eventually reaching the Indian Ocean via the Juba River (IV). These 20 complex patterns of low effective migration resulted in six pockets of relatively high effective migration: in 21 22 western Central African Republic, southern Chad and northern Cameroon (V, Figure 4); the Horn of Africa (VI); South Sudan, northern Uganda and north-western Kenya (VII); southern Kenya, northern and western Tanzania 23 and northern Zambia (VIII); south eastern Tanzania and northern Mozambique (IX); and finally southern Africa 24 roughly south of the Zambezi basin to the Cape of Good Hope (X). We then overlaid the geographic distribution 25 26 of the six putative historic populations in previous analyses and found that the distribution of high and low effective 27 migration areas corresponds largely with major population boundaries (Figure 4).

28

29 Geographic distribution of genome-wide diversity in the black rhinoceros

First, we jointly estimated the effective diversity surface for our georeferenced historic data set using EEMS for 30 an initial idea of the geographic distribution of genome wide diversity (Figure S10). This analysis suggested a 31 32 broad region of high diversity corresponding to East and Central Africa and two regions of low diversity in 33 southern and northwestern Africa. Then, we estimated the genome-wide heterozygosity (GWhet) per sample based 34 on transversion sites. Historically, GWhet was highest in CE (median = 3.28×10^{-4}) and EA (median = 3.25×10^{-4}) ⁴), and lowest in the S population (median = 2.36×10^{-4} , Figure 5A/B). We also estimated levels of inbreeding 35 36 among black rhinoceroses by calculating the average length of homozygous regions, known as runs of 37 homozygosity (RoH), and divided it by the total length of the scaffolds considered (>14Mb, see Variant site

- 1 *identification* in Methods) to obtain individual inbreeding coefficients (F_{RoH} , Figure 5C/D). Among historical 2 samples, we found that F_{RoH} was inversely related to GWhet, being lowest in CE and EA and highest in NW and 3 S. Thus, both GWhet and F_{RoH} peaked in Central and East Africa, at the centre of the species distribution, and 4 decayed toward the northern and southern peripheries (Figure 5 A-D). Both associations were significantly 5 correlated with distance from the individual with the highest GWhet, thus explaining 73% (p < 0.001, Figure S11) 6 and 68% (p < 0.001, Figure S12) of the variation in GWhet and F_{RoH} , respectively. We did not find a significant 7 association between sampling date and either GWhet or F_{RoH} (Figures S13 and S14 respectively).
- 8

9 Modern individuals showed much lower GWhet. Although Kenyan samples (modern CE-EA) showed lower 10 GWhet (median = 2.57×10^{-4}) than their presumed historic sources, EA and CE, these levels were still within the 11 range of some of the historic samples, being comparable to NE and RU. Conversely, modern individuals from 12 South Africa and Namibia (Modern S) featured much lower GWhet than that of any historic populations (Figure 13 5A). However, unlike GWhet, historical F_{RoH} was not significantly different to levels in Modern CE-EA and 14 Modern S (Figure 5C).

15

We explored this breakdown in the relationship between F_{RoH} and GWhet among modern samples by dividing 16 F_{RoH} into three different size classes, with RoH between 1 and 2 Mb equating to inbreeding within the last 43 17 generations (Figure 6A), RoH between 2 and 5 Mb reflects inbreeding within the last 21.5 generations (Figure 6B) 18 and RoH >5Mb equates to inbreeding within the last 8.6 generations (Figure 6C). By assuming a generation time 19 of 24 years (Moodley et al. 2017) we estimated timeframes for historical inbreeding of 517 - 1032 years, 207 -20 516 years and 0 - 206 years for the small, medium, and large F_{RoH} size classes respectively. While southern Africa 21 expectedly showed considerably more recent inbreeding during the colonial period (17th - 20th centuries), along 22 23 with NW and RU; populations in East Africa (CE, EA, and NE) displayed more inbreeding within the two older 24 timeframes.

25

26 The burden of inbreeding

27 The high levels of inbreeding observed in previous analyses necessitated an analysis of the genetic load borne by 28 each population across the species range. We found differences in realised genetic load that is due to homozygous 29 loss of function alleles and the masked genetic load of heterozygous loss of function alleles between populations as well between historic or modern samples (Figure S15). Similar to both GWhet and F_{RoH}, southern Africa (S) 30 appears to suffer the highest burden in both realised and masked genetic load. However, modern S had significantly 31 32 lower realised genetic load than its historic counterpart (Table S6), with large variability between individuals, 33 which could suggest efficient purging of deleterious alleles while masked genetic load between historic and 34 modern S individuals overlapped. On the other hand, we observed no obvious differences in realised genetic load 35 between historic and modern CE and EA populations (Figure S15). The masked genetic load may be somewhat 36 lower in the modern CE/EA population compared to its historic counterparts; however, differences appear only

1 minor.

2

3 **Discussion**

The aim of this study was to characterise the population structure, and the distribution of genomic diversity in the black rhinoceros before its range-wide collapse in the latter half of the 20th Century. Today, the natural populations of black rhinoceros occurring in Kenya, Tanzania, Zimbabwe, Namibia, and South Africa are remnants of a much richer genetic diversity in the recent past. We therefore sourced and analysed whole-genome data from 63 museum specimens representing the continental-scale historic distribution of the species.

9 Historic populations of the black rhinoceros

Our PCA and admixture analyses indicated that historic genomic variation in the black rhinoceros was geographically structured into six major populations (S, RU, CE, EA, NE, and NW), with further substructuring in southern Africa of S into SW and SE/SN subpopulations and in East Africa of CE into MA and ER and EA into NR. We were concerned that gaps in our sampling scheme, e. g. in the Central African Republic, western South Sudan, southern Tanzania, and northern Mozambique, may have contributed to the observed population genomic structure. On the other hand, differentiation into distinct EA and CE populations was observed despite particularly dense sampling in their region of overlap.

17

Further exploration of the nature of population structure showed a significant pattern of isolation by distance, 18 where the genetic distance between pairs of individuals increased as a linear function of geographic distance 19 between their sampling locations. However, we also observed that several of the geographically distant pairwise 20 comparisons (> 2,000 km) were more genetically distinct from each other than would be predicted from the 21 22 distance between their sampling locations (Figure 3A). These outliers suggested genetic discontinuities in parts of 23 the species range. A non-linear decline in D-statistics from the centre of the species range confirmed this 24 observation. Finally, effective migration rates modelled using EEMS defined six regions of high migration that 25 corresponded directly with the six major populations observed in PCA and admixture analyses (Figure 4B). Although this latter analysis was conducted only on our 53 geo-referenced samples, and further data would help 26 define these regions more precisely, taken together we are confident that the genomic variation in black r hinoceros 27 was structured as described above. Our observations of population and subpopulation structure largely corroborate 28 29 the findings of Moodley et al. (2017). However, the increased resolution of our whole genome data set enabled 30 the detection of additional substructuring of populations ER and NR, to the east and north of the rift valley, 31 respectively. Our whole genome data did not retrieve the divergent mtDNA clade WW, despite sampling from 32 west of the Shari-Logone basin in Nigeria. We suggest that WW maternal lineage might be a relic of an ancient 33 migration into West Africa that has become fixed west of the Shari-Logone by genetic drift. This discrepancy, and 34 the high-resolution population structure observed above, highlights the growing necessity for the use of genome-35 scale data to infer intraspecific phylogeography.

1 Evolutionary history of the historic black rhinoceros populations

Central and eastern Africa, dominated by populations EA and CE, appear to have been the hotspots of black
rhinoceros diversity, whereas individual diversity decayed with increasing latitude both northward (populations
NW and NE) and southward (populations RU and S), towards the limits of the species range (Figure 5). Decreasing
genetic diversity from the central parts of a species range is commonly observed in both plant and animal species
(Eckert et al. 2008) and is thought to result from increasing isolation and smaller effective sizes. Genetic diversity
is often, but not always, highest at or near the species origin (Liu et al. 2006), particularly in species exhibiting
significant IBD, as do humans (Manica et al. 2005).

9

We have shown here that up to 67% of the heterogeneity in our whole genome data set reflected IBD and, thus, we propose that Central and/or eastern Africa, east of the Congo basin, as the putative region of origin for the black rhinoceros. This inference is supported by the fossil record, with the earliest emergence of modern *D. bicornis* at Koobi Fora in Kenya, 2.5 million years ago (Ma). The species range then appears to have expanded rapidly, as it appears subsequently at Baard's Quarry in South Africa 2.0 Ma and in the Konso Formation in Ethiopia 1.8 Ma (Geraads 2010).

16

17 Regarding the isolation of the southern African populations, it is known that tectonic upliftment across the Kalahari 18 sands of southern Africa resulted in a drainage depression that gave rise to enormous Lake Palaeo-Makgadikgadi 19 between 1.4 and 0.5 Ma (Moore et al. 2012; Riedel et al. 2014). This event isolated the western and central parts of southern Africa from the basins of the Kavango, Chobe and Upper Zambezi Rivers through what is now central 20 Botswana and ties in with the first split in our phylogeny, separating the ancestors of S and RU from the rest of 21 Africa. The Upper Zambezi was eventually captured 125-150 thousand years ago (Ka) into its present-day course 22 (Moore and Larkin 2001), effectively isolating all of southern Africa (S) from RU and the rest of the continent. 23 We observed the major genetic discontinuity in our range-wide dataset across the axis of the Zambezi River and 24 25 suggest that the series of geological events outlined above may have provided the strongest barrier to gene flow 26 across the historic range of the black rhinoceros. In such a scenario, it seems most plausible that black rhinoceros inhabiting the area west of the central Kalahari were most isolated by Lake Palaeo-Makgadikgadi, potentially 27 explaining why SW is well differentiated from SN/SE subpopulation, both of which would have inhabited the 28 region to the east of the paleo-lake, and with possibly greater access to Central Africa, prior to the capture of the 29 30 Lower Zambezi River. Lake Palaeo-Makgadikgadi is known to have fluctuated greatly in size in the last 50 kya 31 (Riedel et al. 2014), eventually allowing the black rhinoceros to repopulate northern Botswana and southeastern 32 Angola from Zimbabwe (SN).

33

Meanwhile, in the rest of the continent, the black rhinoceros had begun to diverge into populations firstly along the axis of the Albertine and Gregory Rifts, with NW to the west of this system the first to differentiate (Figure S4). CE then became differentiated from EA and NE along the axis of the Tana River, and also potentially via admixture into the latter two populations from NW (Figure S4). Although the upper Tana River altered its course during the middle Pleistocene, it has flowed nevertheless across central Kenya to the Indian Ocean since the
upliftment of the Aberdare Range and Mount Kenya in the late Miocene and Pliocene (Baker et al. 1971;
Veldkamp et al. 2012). Thus, the river and its mountain sources within the Gregory Rift system have likely been
barriers to gene flow throughout the evolutionary history of the black rhinoceros in East Africa. At a finer scale,
the geographic localisation of populations MA, ER and NR to the western, eastern and northem parts of the
Gregory Rift, respectively, provides further evidence that rifting, and upliftment were major drivers of black
rhinoceros differentiation in East Africa.

8

9 Admixture also appears to be a common feature of historical populations. Admixture profiles for $K \ge 2$ (Figure 10 S4) indicate that while barriers to gene flow were important in isolating populations, introgression between populations was common once such barriers were removed. We postulate that admixture, followed by isolation, 11 may have been responsible for the evolution of several populations including RU, EA, MA and ER. Ancient 12 introgression between CE and S explains why RU is closely affiliated to CE on admixture plots (Figures 2F, S4 13 and S6) and yet a sister taxon to S on the phylogenetic tree (Figure 2G). Such gene flow across the lower Zambezi 14 between eastern southern Africa and southeastern parts of central Africa, would only have been possible prior to 15 the capture of the upper Zambezi when the river's flow might not have been as permanent as it is today. Likewise, 16 EA is likely derived through admixture between NW and CE, MA from admixture between CE and RU and ER 17 through admixture between CE and MA. These signatures for admixture may also indicate shared ancestral 18 polymorphisms rather than gene flow, and since neither are accounted for in a bifurcating tree approach, they are 19 20 likely responsible for the low gene and site concordance values observed in our phylogenetic reconstruction. We 21 are hopeful, however, that these evolutionary events can be teased apart in the future, through demographic modelling when greater sample sizes become available. 22

23

In conclusion, although we demonstrate distinct population structure across its range, we also show that the evolutionary history of the black rhinoceros was likely driven not just by allopatric separation of populations owing to the species inability, or reluctance, to cross large and permanent water bodies and mountains, but also by secondary contact followed by isolation, in cases where barriers to gene flow were temporarily removed. Thus, the overlaying of these various evolutionary events upon each other has led to a significant pattern of IBD across the wide sub-Saharan range of the black rhinoceros.

30 Historic levels of inbreeding vary with geography

Assessing the structure of modern samples from Kenya, Namibia and South Africa allowed insight into how population declines have compressed and distributed the remaining historical diversity. The breakdown in relationship between GWhet and F_{RoH} among modern samples is intriguing. While among southern African individuals, where the colonial period began in the 1600s and is known to have heralded the onset of habitat destruction and trophy hunting on a vast scale, it might be expected that historical samples from 1776 (ZA1776.1) and 1845 (ZA1845.1) may already have been subjected to inbreeding at the time of sampling, and so their F_{RoH}

1 values may appear similar to levels in modern samples from the same region. However, in East Africa, similarly 2 scaled interventions by Europeans began much later, in the mid to late 1800s, and so most historical samples in 3 our data set were expected to have significantly lower levels of RoH. By dividing RoH into size (and hence time) 4 classes, we show that East African black rhinoceros (CE, EA and NE), while possessing fe wer large RoH tracts 5 than S, RU and NW, still contained appreciable levels of colonial-period inbreeding (Figure 6C), showing for the 6 first time, the negative genetic consequences of the ubiquitous European hunting safari on black rhinoceros 7 diversity in East Africa. However, the same three East African populations had significantly more RoHat medium 8 and small size classes (Table S1 B). This provides evidence of precolonial inbreeding among black rhinoceros in 9 East Africa, whereas populations outside this region, in West Africa, southern east Africa and southern Africa 10 show inbreeding mainly during the colonial period. A similar result was shown for white rhinoceros, were effective population sizes among southern white rhinoceros of southern Africa were lowest during the colonial period, 11 12 whereas values for the northern white rhinoceros were lowest during Bantu migrations into East Africa (Moodley et al. 2018). Thus, independent genetic data from both African rhinoceros species point to geographically distinct 13 patterns of inbreeding between southern and East Africa, suggesting that anthropogenic pressures on African 14 rhinoceros date back to antiquity, and may have been, as it is today, associated with rhinoceros horn. This view is 15 corroborated by the fact that rhinoceros horn, and other wildlife products from East Africa, were already being 16 traded along the Arabian coast and further east by 100 AD (Boeyens and van der Ryst 2014). 17

18 Conservation implications

Our whole genome dataset provides the first resolution of nuclear DNA populations NW, NE, RU, and MA, which 19 were previously only suggested by mtDNA, plus two entirely unknown populations ER and NR. Moreover, our 20 genome scale, geo-referenced data set allowed the more precise localisation of all black rhinoceros populations 21 22 and subpopulations across the species range than was previously possible with spatial modelling of low-resolution traditional markers (Moodley et al. 2017). One such example occurred in East Africa, where variation neither at 23 24 microsatellites nor at mtDNA was able to resolve the geographic ranges of populations EA and CE. Here, we show that the distribution of EA is clearly distinguishable from CE, with the former ranging in suitable habitat between 25 26 the Albertine and Gregory Rifts and the latter distributed from about the Tana River south to the Zambezi River, with a zone of secondary contact between EA and CE in southern Kenya. Our genome data also identified distinct 27 28 black rhinoceros subpopulations MA, ER and NR that we localised to different regions within the rift valley. Thus, 29 the additional structure and better geographic localisation of populations offered by whole genome data have major implications for conservation-oriented management. 30

31

Unfortunately, both NW and NE have been extirpated, with no known record of animals from those populations ever successfully contributing to *ex-situ* populations. On the other hand, confirmation of the existence of RU places enormous conservation value on any of its remaining individuals in the wild. Its historic range, covering the eastern part of Central Africa from the Zambezi River in the south to the Rufiji in the north, contains only two possible options for the persistence of RU individuals: Selous Game Reserve in Tanzania and Niassa Game Reserve in Mozambique. With no recent reports of black rhinoceros activity in either reserve, and with local authorities
 incapable of providing the necessary protection, the future of RU, and its unique portion of black rhinoceros
 diversity looks bleak. Similarly bleak prognosis can be made about the existence of ER and NR among modern
 populations.

5

6 The modern samples highlight the devastating effect of population contractions and subsequent genetic drift. This
7 observation was shown to be worst among modern southern African individuals, which featured the lowest
8 heterozygosity and highest inbreeding across all populations, descending from a limited number of founding
9 individuals in Damaraland and Kaokoland, Namibia (SW) (Endangered Wildlife Trust 1984), Zululand, South
10 Africa (SE), and the Zambezi Valley, Zimbabwe (SN; Emslie and Brooks 1999).

11

12 In Kenya, despite sustaining precolonial inbreeding associated with Bantu migrations, colonial inbreeding associated with European hunting and finally the heavy population contractions from the 1970s to the 1990s, 13 modern Kenvan black rhinoceros still maintain much higher levels of present-day variation than modern southern 14 African populations (Figure 5). At one stage, the plight of the Kenyan black rhinoceros was so serious that local 15 authorities located, caught and translocated the last animals from the dwindling populations scattered across that 16 country, in a desperate effort to consolidate the national metapopulation into intensive protection zones (IPZs). In 17 the absence of genetic knowledge at that time, the origin of each animal was not considered, and so EA, NR, CE 18 and ER individuals were inadvertently placed within the same IPZs. However, two IPZs in Kenya never received 19 introductions from elsewhere, and these were the Maasai Mara Game Reserve and Chyulu National Park. 20

21

30

These management decisions have resulted in the admixture of EA and CE in much of the present day Kenyan 22 23 metapopulation, as is clear from the intermediate PC space occupied by most modern day Kenyan samples (Figure 24 2C), and in contrast to separately managed southern African populations (Figure 2E), where modern and historic 25 samples cluster together. In our data set, individuals with highly admixed EA/CE profiles were typically from 26 Nairobi National Park and the Ol Pejeta Conservancy (Figure S6). While it is possible that typically NR and ER may also have contributed to the diversity of the present-day Kenyan metapopulation, our restricted modern 27 sample from Kenya did not allow for their detection. The conservation benefit of the consolidation of the Kenyan 28 29 metapopulation was thus the maintenance of high genetic diversity in the face of population collapse.

On the other hand, our modern samples from the Maasai Mara, possess lower genomic diversity compared to other 31 32 Kenyan populations but, because no translocations ever entered this IPZ, they reveal the original mix of population ancestries that would have been present in Maasailand in historical times. Therefore, all three genomes sampled 33 34 in the Maasai Mara were typically of Maasailand (MA) ancestry, and probably represent the last place in Africa, together with the adjoining Serengeti, where the MA population still exists. It is unlikely that MA would have 35 36 survived in the Massai Mara had this reserve been part of the original translocation plans to protect the Kenyan 37 black rhinoceros, and so our results vindicate the original decision to manage this reserve separately from others 38 in Kenya.

2 A similar situation may exist among the non-admixed black rhinoceros population of Chyulu National Park. 3 Although this population was reduced to only two individuals in 1992, it had grown to 21 by 2011 (Muya et al. 2011). Chyulu is in southern Kenya and importantly, to the east of the rift valley. It may therefore still harbour 4 5 individuals with ER ancestry, although the national park has never been sampled. From a conservation perspective, 6 this possibility alone elevates Chyulu National Park to a similar level to that of the Maasai Mara as it may be the 7 last place in Africa where ER might exist. Another interesting possibility is that ER and CE may exist in ex situ 8 black rhinoceros populations that were removed from southern Kenya, east of the rift valley during the 1960s 9 (Moodley et al. 2017). These ex-situ populations can be found at Thabo Tholo reserve in South Africa, where 10 unfortunately, many have been admixed with S individuals and thus unsuitable for reintroduction anywhere in East Africa. Another possibility for the existence of ER and CE is in European and American zoos, particularly 11 12 Dvůr Králové Zoo in the Czech Republic, whose black rhinoceros collection stems directly from Tsavo National 13 Park, also to the east of the rift valley (Moodley et al. 2017).

14

1

Based on these results, we suggest strictly separate management for the Maasai Mara-Serengeti and Chyulu 15 National Park from each other and the rest of the Kenyan metapopulation. We suggest local authorities step up 16 measures to genetically profile all remaining black rhinoceroses in Kenya, particularly for those populations with 17 little or no genetic data. The overarching goal for the long-term management of the Kenyan black rhinoceros 18 would be to maintain MA in the Maasai Mara, potentially ER and CE in Chyulu, and EA/CE within the remaining 19 metapopulation, with regular monitoring to sustain levels of diversity, attenuate genetic drift and limit inbreeding. 20 Similarly, but more urgently, we recommend that authorities in Tanzania obtain genetic data for all their remaining 21 black rhinoceros, with their top priorities to maintain both RU, MA, ER and CE populations, wherever they might 22 23 still occur in that country.

24

In southern Africa, our results confirmed previous findings, and we therefore recommend a continu ation of the 25 26 current management scheme, where SW (the Namibian black rhinoceros) is managed separately from subpopulation SN/SE. Our results also confirm the close relationship between SN and SE, which were previously 27 managed separately. We suggest, as did Moodley et al (2017), that new reserves established anywhere in eastern 28 29 southern Africa from the Cape to the Zambezi consider founders from both SN and SE when available. As both 30 the realised and masked genetic loads were highest among southern African black rhinoceros, we recommend measures to avoid further inbreeding, such as the movement of males between reserves and population monitoring 31 32 using a studbook, be implemented in all facilities with small populations, whether wild or captive. We also caution 33 that although individual numbers are highest in southern Africa, these populations represent but a small fraction 34 of the remaining species diversity and a conservation management focus on maintaining as many different genetic 35 populations is now required, rather than simply increasing numbers and growth rates of southern African black 36 rhinoceros.

37

38 Beyond these conclusions, having genome-wide data available opens promising new avenues for conservation-

1 related research on the black rhinoceros. Our map of black rhinoceros genomic diversity could be leveraged to 2 develop more sophisticated molecular tools to identify the provenance of black rhinoceros material seized from 3 the illegal market. Also, with genomic information we could venture into the potential phenotypic effects of the 4 intra-species diversity observed in order to guide management actions. For instance, gaining insight into local adaptation, inbreeding and outbreeding depression might greatly enhance the success of breeding programs. 5 6 Overall, our results support and highlight the importance of improving the resolution of traditional molecular 7 markers by carrying out population level, whole-genome studies, and by sampling widely across the species range 8 to better understand population structure and evolutionary history, and ultimately, to better inform conservation 9 management.

10

11 Acknowledgements

12 This work was supported by ERC Consolidator Grant 681396 'Extinction Genomics' to M.T.P.G. and by EMBO Short-Term Fellowship 7578 to F.S.B. The authors would like to acknowledge support from Science for Life 13 Laboratory, the National Genomics Infrastructure (NGI), Sweden, the Knut and Alice Wallenberg Foundation and 14 UPPMAX for providing assistance in massively parallel DNA sequencing and computational infrastructure. TMB 15 is supported by funding from the European Research Council (ERC) under the European Union's Horizon 2020 16 research and innovation programme (grant agreement No. 864203), BFU2017-86471-P (MINECO/FEDER, UE), 17 "Unidad de Excelencia María de Maeztu", funded by the AEI (CEX2018-000792-M), NIH 1R01HG010898-18 19 01A1. YM acknowledges support from the National Research Foundation of the Republic of South Africa. 20

The authors are very grateful to all the museums who contributed samples to this study: the Natural History Museum London, the Museum of Natural History Berlin, the Powell-Cotton Museum, the Natural History Museum Vienna, the Natural History Museum at the National Museum Praha, the Natural History Museum of Zimbabwe, the Swedish Museum of Natural History, the Royal Museum for Central Africa Tervuren, the Senckenberg Museum Frankfurt, the United States National Museum Washington (or Smithsonian Arts and Industries Building) and the Bavarian State Collection of Zoology.

27

30

The authors would also like to thank Professor Alfred L. Roca (Department of Animal Sciences, University of
Illinois Urbana-Champaign) for insightful comments on the preliminary results of this manuscript.

Lastly, it is with great sadness that we acknowledge the death of our friend, colleague, mentor and co-author Michael W. Bruford, who succumbed to illness during the resubmission phase of this manuscript. He will be sorely missed far beyond the bounds of just the conservation genetics community.

34

35 Data Availability

36 The sequencing data underlying this article is available on SRA under BioProject Number PRJNA1002571.

1 Methods

2 Whole-genome data generation

3 Our historic sample collection included material obtained from 71 museum specimens. Collection dates ranged 4 between 1775 and 1981, with the oldest sample a bonafide representative of the Cape rhinoceros (D. b. bicornis), 5 which was thought to be extinct. All samples consisted of keratinous material (pieces of skin, horn powder, or 6 hairs), except for ZA1845.1, which was a piece of bone from a skull, and ZA1775.1, which was a molar tooth. 7 Samples from historic specimens were stored and processed in facilities dedicated to ancient DNA work at the 8 Swedish Museum of Natural History (Stockholm), and the Natural History Museum of Denmark (Copenhagen). We followed Sánchez-Barreiro et al. (2021) for keratinous tissue processing. The skin pieces were manually cut 9 10 and then hydrated for 2–3 h at 4°C in 0.5–1 mL of molecular biology grade water. The tissue was then briefly washed with 0.5 mL of a 1% bleach solution, followed by two rinsing steps with molecular biology grade water" 11 12 (Sánchez-Barreiro et al., 2021). Bone material was crushed with a small hammer, and small pieces amounting to 150-200 mg were used for extraction after a brief washing with a 1% bleach solution, and two rounds of rinsing 13 with molecular biology grade water. Our collection also included 27 modern samples in the form of keratinous 14 material either preserved in ethanol or dry. Dry samples were hydrated with molecular biology grade water prior 15 to manipulation, and then each piece of skin was cut with a disposable scalpel. For extraction, 20 mg of material 16 17 was used.

We extracted DNA from the historic keratinous samples with the DNeasy Blood and Tissue Kit (Qiagen), but 18 introducing two modifications to the manufacturer's guidelines, as indicated in Sánchez-Barreiro et al. (2021): on 19 one hand, "adding of DTT (dithiothreitol) 1 M to a final concentration of 40 mM to the lysis buffer", and also "the 20 substitution of the purification columns in the kit by MinElute silica columns (Qiagen) to favour retention of small 21 fragments". DNA extraction from the bone and the tooth samples was carried out following (Gilbert et al. 2007) 22 23 with the modifications detailed in Dabney et al. (2013) to enhance the retrieval of small DNA molecules. We assessed the concentration and fragment size distribution in each extract using a TapeStation 2200 (Agilent). 24 25 Extraction of DNA from the modern samples was carried out with the KingFisher[™] Duo Prime instrument and 26 its associated Cell and Tissue DNA Kit, following the manufacturer's guidelines. Concentration of DNA extracts 27 was measured with a Thermo ScientificTM Qubit dsDNA high-sensitivity (HS) assay. A 20 µL aliquot of each 28 extract was fragmented in a Covaris® focused-ultrasonicator with a customised program to reduce fragment length to ~400 bp. Size distribution upon fragmentation was assessed with a TapeStation 2200 (Agilent, Santa Clara, 29 CA). 30

Sequencing library preparation followed the procedure described in (Sánchez-Barreiro et al. 2021), using the BEST protocol (Carøe et al. 2018). We used 100 ng of extracted DNA to which we ligated adapter sequences compatible with BGISEQ 500 sequencing (Mak et al. 2017). Libraries were PCR amplified and single-indexed following strictly the protocol described in Sánchez-Barreiro et al. (2021). Resulting indexed libraries were distributed in pools containing equimolar proportions of eight indexed libraries each. Each of these pools was

- 36 given one lane of BGISEQ 500 PE150 sequencing.
- For samples ZA1, ZA2 and NA1, sequencing libraries were built using the Illumina® TruSeq® Nano DNA

1 Library Prep Kit for NeoPrep[™] on DNA inserts that were 350 bp in length and following the manufacturer's

2 guidelines. Libraries were then sequenced on an Illumina® HiSeq X platform, giving 0.5 lanes per sample in

3 PE150 mode.

4 Bioinformatic processing of raw data

5 Quality assessment and mapping of DNA sequence data

6 We generated shotgun sequencing data for a total of 98 black rhinoceros samples, 71 historic and 27 modern. We 7 conducted a quality check per sample with fastqc v0.11.7 (Andrews 2010). Subsequently, we ran the pipeline 8 PALEOMIX v1.2.13.2 (Schubert et al. 2014) on each sample separately to: remove sequencing adapters and 9 exclude reads shorter than 25 bp with AdapterRemoval v2.2.2 (Schubert et al. 2016); align the raw reads against the Diceros bicomis assembly ASM1363453v1 (Genbank Assembly Accession: GCA 013634535.1; Moodley et 10 al. 2020) using bwa v0.7.16a and its backtrack algorithm (Li and Durbin 2009) setting minimum base quality 11 filtering to zero to maximise reads retained; filter out duplicates with Picard MarkDuplicates (Broad Institute 12 2019); calculate the level of ancient DNA damage with mapDamage v2.0.6 (Jónsson et al. 2013). From the total 13 98 samples, 8 samples were excluded from whole-genome analyses due to low depth of coverage (<1x; see Table 14 S1) or systematic failure to align against the whole-genome assembly. The resulting 90 aligned genomes were 15 16 divided into 63 historic and 27 modern.

17 Variant site identification

To optimise computational memory usage and omit potentially poorly assembled regions of the reference 18 assembly, we restricted variant site finding to scaffolds >14 Mbp (n = 47), which represent 72.83% of the total 19 length of the assembly. We verified that none of these scaffolds belonged to sex chromosomes by evaluating if 20 male samples showed a 0.5x normalised depth of coverage, indicative of X chromosome regions (Figure S3). We 21 identified biallelic variant sites that were transversions and computed their genotype likelihoods using the GATK 22 23 genotype likelihood model (-GL 2) within ANGSD v0.921 (Korneliussen et al. 2014). Transitions were excluded with the -rmTrans option, and the minimum number of individuals in which a variant site must be present (-24 25 minInd) was 95%. Minimum and maximum global depth per site were based on a global depth assessment with 26 ANGSD -doDepth: 500 and 1,500 respectively when including 63 or more genomes; 200 and 1,500 when 27 including fewer than 63 genomes. Additionally, the following quality filtering and output choice parameters were 28 set: -remove bads 1 -uniqueOnly 1 -bag 1 -C 50 -minMapO 30 -minO 20 -doCounts 1 -GL 2 -doGlf 2 -29 doMajorMinor 1 -doMaf 1 -doHWE 1 -dosnpstat 1 -HWE pval 1e-2 -SNP pval 1e-6.

30 Statistical analyses of genomic data

31 Relatedness test

32 We ran a pairwise analysis of relatedness based on genotype likelihoods with ngsRelate v2 (Hanghøj et al. 2019).

33 The computation of this panel of genotype likelihoods followed the procedure detailed above, except for the



2 genotype likelihoods and allele frequencies were reformatted with commands in bash language to match the input

3 requirements of ngsRelate v2. As per (Waples et al. 2019), the degree of relatedness between each pair of samples

- 4 was assessed qualitatively based on the relative values of coefficient of relatedness R1 versus coefficients KING
 5 and R0.
- We found 11 pairs of individuals showing a relatedness signal among the modern samples (Figure S1). Seven of
 those samples were therefore excluded from the analyses of population structure: MA1, MA2, MA5, MA7, OP10,
 OP11, ZA2. As a criterion to exclude samples from a related pair, the sample of lowest depth of coverage was
 discarded.
- **10** Principal Component Analysis (PCA)
- We used PCAngsd 0.973 (Meisner and Albrechtsen 2018) to compute covariance matrices from genotype likelihoods for different sets of samples: all historic genomes, those north and south of the Zambezi River separately, and the latter plus the unrelated modern genomes stemming from those respective regions. Standard packages in R v3.4.4 (R Core Team 2022) were used for decomposition of each matrix in eigenvectors and eigenvalues, and *ggplot2* (Wickham 2016) for visualisation of principal components (PCs).

16 Admixture

- Assessment of admixture proportions across individuals was conducted with NGSadmix v 32 (Skotte et al. 2013). We used the genotype likelihoods of transversion variant sites for the 63 historic genomes as input. Values of ancestral clusters, K, ranged between two and ten, and for each value of K, we ran NGSadmix 100 times. We repeated the analyses with the inclusion of the modern individuals for K values 2-10. For each value of K, the run of highest log-likelihood was chosen for visualisation with the software Pong (Behr et al. 2016)(Figure S4 and S5). We used EvalAdmix (Garcia-Erill and Albrechtsen 2020) to evaluate the goodness of fit of the clustering for each K value (Figure S6 and S7).
- 24 Phylogenetic tree

We selected a single individual per population (CEN, CES, NE, NW, RU, S) that showed the low levels of mixed 25 ancestry based on K=6 in the admixture analysis; KE1911.1, KE1933.1, TZ1910.1, TD1925.2, SO1896.2, 26 27 ZW1880.1. We additionally mapped a white rhinoceros individual (P9109_108) to the black rhinoceros genome 28 to act as an outgroup. The white rhinoceros individual was mapped to the black rhinoceros reference genome using 29 PALEOMIX, following the same protocol described above for the black rhinoceros data. We generated consensus 30 fasta files from each of the individuals using ANGSD and a consensus haploid call (-doFasta 2) and the following 31 filters: -remove bads 1 -uniqueOnly 1 -minMapQ 30 -minQ 20 -setmindepthind 5. We limited this to scaffolds 32 >14Mb. We generated a bed file containing sliding windows of 20kb in size with 1Mb slides using bedtools v2.29.1 (Quinlan and Hall 2010) and extracted each window from the individual specific consensus file using 33 34 SAMtools. We built a phylogenetic tree for each window (gene tree) using IQ-tree v2.2.0.3 (Minh et al. 2020) 35 with the GTR substitution model + six gamma distribution rate categories (R6) and 1,000 bootstrap replicates. We 36 also concatenated all windows into a single sequence and built a phylogenetic tree using IQ-tree. We calculated

gene concordance factors (percentage of gene trees supporting a given node), and site concordance factors
(percentage of sites supporting a given node) based on the topology from the concatenated data and the individual
gene trees in IQ-tree (--gcf and --scf). We dated the concatenated tree using MCMCtree from the PAML package
(Yang 2007) and specified a root age (split between black and white rhinoceros) between 5.3 and 7.3 Ma. This
range is based on records of *Diceros* in upper Miocene deposits (> 5.3 Ma) at Lothagam (Kenya, 6.54-5.2 Ma;

6 (Brown and McDougall 2011)) and Albertine (Uganda, 7.25-5.3 Ma; Pickford et al. 1993).

7 Factors influencing genetic distance

8 To perform a Mantel test for isolation by distance (IBD) we generated two distance matrices. One based on genetic distance, and one based on geographic distance. We calculated the genome-wide pairwise distance between either 9 all 53 georeferenced historical black rhinoceros or 52 (we excluded one southern African individual (ZA1775.1) 10 11 due to elevated putative genetic distances caused by low coverage data (1.27x) using ANGSD with a consensus base call (-doIBS 2) and the following parameters: -rmtrans 1 -minind 53 -remove bads 1 -uniqueOnly 1 -12 minMapO 30 -minO 20 -GL 1 -doMajorMinor 1 -minMinor 0 -makeMatrix 1. Similar to the other analyses we 13 limited our analysis to scaffolds >14Mb in length. We generated the geographic distance matrix for the same 14 15 individuals using their GPS coordinates (Table S1) and R using the geodist library. We ran the Mantel test in R 16 specifying the two distance matrices as input and 9,999 permutations. We also performed a regression test by 17 comparing pairwise differences between dates and pairwise genetic distance of the same individuals to assess 18 whether there was a temporal factor driving the genetic differences between samples. The correlation coefficient was calculated using R v4.2.1 (R Core Team 2022). 19

20 D-statistics

To estimate the relatedness of the individuals found in the central range (CE or EA) of the species to either the 21 southern (S) or northern (NW) populations, we used D-statistics in ANGSD. We used a random base call (-22 23 doabbababa1), specified the white rhinoceros (Biosample accession: SAMEA8896056) as the outgroup, only used 24 scaffolds >14Mb in length, excluded repeat regions, and chose the following parameters: -remove bads 1 -25 uniqueOnly1-baq1-C50-minMapQ30-minQ20-setMaxDepth1500-setMinDepth500-rmTrans1. The output 26 was parsed through the jackKnife.R script, which is part of the ANGSD toolsuite, to make it into a more readable 27 format. ANGSD calculates D-statistics for all possible triplet combinations. However, we only extracted 28 comparisons following the defined topology of (((S,NW), central population), Outgroup). Based on this topology, 29 a negative D-score would indicate a closer relationship to the S population, whereas a positive D-score would 30 indicate a closer relationship to the NW population. As we have multiple individuals from S and NW, we took the 31 average of all possible combinations of S/NW.

32 Estimation of effective migration and diversity surfaces with EEMS

We employed EEMS (Petkova et al. 2016) to link genetic and geographic data and estimated the effective migration and diversity surfaces along the black rhinoceros range of distribution using 53 georeferenced historic

35 genomes. As input, EEMS takes a pairwise distance matrix which we calculated with PLINK using an input file

36 generated using ANGSD (-doplink 2) across the 47 largest scaffolds of the assembly and the following parameters:

-rmtrans 1 -minind 51 -remove_bads 1 -uniqueOnly 1 -minMapQ 30 -minQ 20 -GL 1 -doMajorMinor 1 -doPlink
2 -doGeno -4 -doPost 1 -postCutoff 0.95 -SNP_pval 1e-6 -doMaf 1 -minMaf 0.05. Using PLINKv1.90b6.2, we
converted the resultant tped and tfam to map/pedfiles using --recode and then converted those to bed/fam files
using --make-bed. From the bed file we generated a distance matrix as input for EEMS using bed2diffs_v1, part
of the EEMS toolsuite. The matrix was fed as input to EEMS with an MCMC chain of 2,000,000 iterations and
assuming 1,000 underlying demes (a specification of grid size). The geographic area of interest was outlined by
hand with the online tool Google Maps API v3 Tool (Scharning). Visualisation of the estimated migration (m) and

8 effective diversity (q) surfaces was conducted in R v3.4.4 (R Core Team 2022).

9 Metrics of individual genomic diversity

We estimated the genome-wide heterozygosity of each genome, based on transversion biallelic sites within the 10 11 scaffolds >14 Mbp, following strictly the approach described in Sánchez-Barreiro et al. (2021). Briefly, for each sample we first calculated the site allele frequency likelihood of there being zero, one or two alternative alleles 12 with the -doSaf 1 option of ANGSD (Korneliussen et al. 2014) and the folded option (-fold 1). Both the reference 13 (-ref) and the ancestral (-anc) genome used were the black rhinoceros assembly. We only included transversion 14 15 sites (-noTrans 1), and sites of depth of coverage of at least 5x (-setMinDepth 5). Identical quality filtering 16 parameters as for computing genotype likelihoods were set. Then we used RealSFS, within ANGSD, to compute 17 the folded site frequency spectrum (SFS) for each sample using the output of the previous step. To investigate the 18 variance of heterozygosity across the genome we calculated the SFS in 10Mb windows of covered bases (-nSites). The count of heterozygous sites was divided by the total count of sites to obtain the individual estimate of genome-19 20 wide heterozygosity.

Runs of homozygosity (RoH) were also estimated for each genome in our data set with >5x coverage using PLINK 21 based on the approach used by Foote et al (2021). We generated a PLINK file from the scaffolds >14Mb in length 22 from all individuals using ANGSD (-doPlink 2) and the following parameters: -rmtrans 1 -minind 83 -23 remove_bads 1 -uniqueOnly 1 -minMapQ 30 -minQ 20 -GL 1 -doGlf 2 -doMajorMinor 1 -doPlink 2 -doGeno -4 24 25 -doPost 1 -postCutoff 0.95 -SNP pval 1e-6 -doMaf 1 -minMaf 0.05. We ran the resultant PLINK file in PLINK to calculate the RoH using the following parameters: --homozyg-snp 50 --homozyg-kb 1000 --homozyg-density 26 50 -- homozyg-gap 1000 -- homozyg-window-snp 50 -- homozyg-window-het 5 -- homozyg-window-missing 5 --27 homozyg-window-threshold 0.05 -- allow-extra-chr. Individual inbreeding coefficients (F_{RoH}) were calculated by 28 29 dividing the total length within RoH > 1Mbp by the total number of bp found in the scaffolds > 14Mb in length (1,698,121,211 bp). 30

31

We also filtered the output into three different RoH categories: 1Mb - 2Mb, 2Mb - 5Mb and >5Mb. We estimated the number of generations since inbreeding occurred using the calculation g=100/(2rL; Kardos et al. 2018), where r = recombination rate, L = length of RoH in Mb, and g= number of generations. As genome-wide recombinationrates for black rhinoceros are unavailable, we present results based on the horse (*Equus Caballus*, 1.16cM per Mb;Beeson et al. 2020). Given this calculation, RoH >1Mb equate to inbreeding occuring within the last 43 1 generations, RoH >2Mb equate to inbreeding occuring within the last 21.5 generations, and RoH >5Mb equates

2 to inbreeding occuring within the last 8.6 generations.

3

We performed regressions of the original sampling date of the individual and GWhet and F_{RoH} as well as geographic distance to the central population and GWhet and F_{RoH} . For the latter, we picked distance to the individual with the highest mean GWhet with GPS coordinates (TZ1910.2) as the central point of the species and calculated distance from that individual taken from the geographic distance matrix calculated above. We limited our analyses to the 52 georeferenced historic black rhinoceros individuals with the exclusion of one southern African individual (ZA1775.1) due to low coverage (1.27x). The correlation coefficients were calculated using R v4.2.1 (R Core Team 2022).

11 Genetic load

Genetic load was estimated to explore the potential consequence of genomic erosion for each individual with 12 sequencing depth >5x following the approach described in Sánchez-Barreiro et al. (2021). Bcftools v1.15 13 (Danecek et al. 2021) was used to call genotypes within scaffolds >14Mb in length. We masked the individual 14 15 genotype as missing for samples with sequencing depth lower than 5x or samples showing heterozygous genotype 16 with either allele having less than 3 reads of coverage. We excluded transition sites, and SNPs with fewer than 2 allele counts, or having over 20% missing information. We used SnpEff v5.1d to annotate the function of each 17 18 variation. For simplicity, we considered the major allele of our black rhinoceros samples as the ancestral state. We then counted the total number of non-synonymous and loss-of-function homozygous, and heterozygous sites 19 separately for each sample to estimate the realised and masked genetic load (Bertorelle et al. 2022). 20

21 Visualisations

All visualisations were produced in R v3.4.4 (R Core Team 2022) using standard packages and *ggplot2* (Wickham
2016). Visualisation of maps and geographical data required the packages *maps* (Richard A. Becker et al. 2018), *mapdata* (Richard A. Becker and by Ray Brownrigg. 2018), *maptools* (Bivand and Lewin-Koh 2019), *rgdal*(Bivand et al. 2019) and *sp* (Pebesma, E.J., R.S. Bivand 2005; Roger S. Bivand, Edzer Pebesma, Virgilio GomezRubio 2013).

27 **References**

Anderson-Lederer RM, Linklater WL, Ritchie PA. 2012. Limited mitochondrial DNA variation within South
 Africa's black rhino (Diceros bicornis minor) population and implications for management. *Afr. J. Ecol.* 50:404–
 413.

Andrews S. 2010. FastQC: a quality control tool for high throughput sequence data.

Baker BH, Williams LAJ, Miller JA, Fitch FJ. 1971. Sequence and geochronology of the Kenya rift volcanics.
 Tectonophysics 11:191–215.

34 Barbosa S, Mestre F, White TA, Paupério J, Alves PC, Searle JB. 2018. Integrative approaches to guide

35 conservation decisions: Using genomics to define conservation units and functional corridors. *Mol. Ecol.*

36 27:3452–3465.

- Beeson SK, Mickelson JR, McCue ME. 2020. Equine recombination map updated to EquCab3.0. *Anim. Genet.* 51:341–342.
- Behr AA, Liu KZ, Liu-Fang G, Nakka P, Ramachandran S. 2016. pong: fast analysis and visualization of latent
 clusters in population genetic data. *Bioinformatics* 32:2817–2823.
- Bertorelle G, Raffini F, Bosse M, Bortoluzzi C, Iannucci A, Trucchi E, Morales HE, van Oosterhout C. 2022.
 Genetic load: genomic estimates and applications in non-model animals. *Nat. Rev. Genet.* 23:492–503.
- 7 Bivand R, Keitt T, Rowlingson B. 2019. rgdal: Bindings for the "Geospatial" Data Abstraction Library.
 8 Available from: https://CRAN.R-project.org/package=rgdal
- Bivand R, Lewin-Koh N. 2019. maptools: Tools for Handling Spatial Objects. Available from: https://CRAN.R project.org/package=maptools
- 11 Boeyens, JC and Van der Ryst, MM, 2014. The cultural and symbolic significance of the African rhinoceros: a
- 12 review of the traditional beliefs, perceptions and practices of agropastoralist societies in southern 13 Africa Southern African Humanities 26(1) pp 21 55
- **13** Africa. *Southern African Humanities*, *26*(1), pp.21-55.
- 14 Broad Institute. 2019. Picard Toolkit. Available from: https://broadinstitute.github.io/picard/
- Brown FH, McDougall I. 2011. Geochronology of the Turkana depression of northern Kenya and southern
 Ethiopia. *Evol. Anthropol.* 20:217–227.
- Carøe C, Gopalakrishnan S, Vinner L, Mak SST, Sinding MHS, Samaniego JA, Wales N, Sicheritz-Pontén T,
 Gilbert MTP. 2018. Single-tube library preparation for degraded DNA. *Methods Ecol. Evol.* 9:410–419.
- Coates DJ, Byrne M, Moritz C. 2018. Genetic Diversity and Conservation Units: Dealing With the SpeciesPopulation Continuum in the Age of Genomics. *Frontiers in Ecology and Evolution* 6:165.
- 21 Dabney J, Knapp M, Glocke I, Gansauge M-T, Weihmann A, Nickel B, Valdiosera C, García N, Pääbo S,
- 22 Arsuaga J-L, et al. 2013. Complete mitochondrial genome sequence of a Middle Pleistocene cave bear
- reconstructed from ultrashort DNA fragments. *Proc. Natl. Acad. Sci. U. S. A.* 110:15758–15763.
- 24 Danecek P, Bonfield JK, Liddle J, Marshall J, Ohan V, Pollard MO, Whitwham A, Keane T, McCarthy SA,
- 25 Davies RM, et al. 2021. Twelve years of SAMtools and BCFtools. *Gigascience* [Internet] 10. Available from:
- 26 http://dx.doi.org/10.1093/gigascience/giab008
- Du Toit, R, 1987. African rhino systematics the existing basis for subspecies classification of black and
 white rhinos. *Pachyderm* 9, 3–7.
- Eckert, C. G., K. E. Samis, and S. C. Lougheed. 2008. "Genetic Variation across Species' Geographical Ranges:
 The Central-Marginal Hypothesis and beyond." *Molecular Ecology* 17 (5): 1170–88.
- 31 Emslie R. 2020. IUCN Red List of Threatened Species: Black Rhino. IUCN Available from:
- 32 https://www.iucnredlist.org/species/6557/152728945#assessment-information
- 33 Emslie R, Brooks M. 1999. African rhino: status survey and conservation action plan. (IUCN/SSC African
- 34 Rhino Specialist Group, editor.). Gland, Switzerland and Cambridge, UK: IUCN
- Endangered Wildlife Trust. 1984. Rhinoceros in South and South West Africa.
- 36 Ferreira, S.M., Ellis, S., Burgess, G., Baruch-Mordo, S., Talukdar, B. & Knight, M.H. 2022. The African and
- 37 Asian Rhinoceroses Status, Conservation and Trade: A report from the IUCN Species Survival
- 38 Commission (IUCN/SSC) African and Asian Rhino Specialist Groups and TRAFFIC to the CITES
- 39 Secretariat pursuant to Resolution Conf. 9.14 (Rev. CoP15). CoP19 Doc. 75 (Rev. 1), CITES
- 40 Secretariat, Geneva, Switzerland.
- 41

- 1 Foote AD, Hooper R, Alexander A, Baird RW, Baker CS, Ballance L, Barlow J, Brownlow A, Collins T,
- 2 Constantine R, et al. 2021. Runs of homozygosity in killer whale genomes provide a global record of
- 3 demographic histories. *Mol. Ecol.* 30:6162–6177.
- Garcia-Erill G, Albrechtsen A. 2020. Evaluation of model fit of inferred admixture proportions. *Mol. Ecol. Resour.* 20:936–949.
- 6 Geraads D. 2010. Rhinocerotidae. In: Cenozoic Mammals of Africa. University of California Press.
- 7 Gilbert MTP, Haselkorn T, Bunce M, Sanchez JJ, Lucas SB, Jewell LD, Van Marck E, Worobey M. 2007. The
- 8 isolation of nucleic acids from fixed, paraffin-embedded tissues-which methods are useful when? *PLoS One*
- 9 2:e537.
- Groves, C. P. Geographic variation in the black rhinoceros *Diceros bicornis* (L., 1758). 1967. Z. Säugetierkd.
 32, 267–276).
- 12
- 13 Hanghøj K, Moltke I, Andersen PA, Manica A, Korneliussen TS. 2019. Fast and accurate relatedness estimation
- from high-throughput sequencing data in the presence of inbreeding. *Gigascience* [Internet] 8. Available from:
 http://dx.doi.org/10.1093/gigascience/giz034
- 16 Harley EH, Baumgarten I, Cunningham J, O'Ryan C. 2005. Genetic variation and population structure in
- 17 remnant populations of black rhinoceros, Diceros bicornis, in Africa. *Mol. Ecol.* 14:2981–2990.
- Hohenlohe PA, Funk WC, Rajora OP. 2021. Population genomics for wildlife conservation and management.
 Mol. Ecol. 30:62–82.
- Jónsson H, Ginolhac A, Schubert M, Johnson PLF, Orlando L. 2013. mapDamage2.0: fast approximate
 Bayesian estimates of ancient DNA damage parameters. *Bioinformatics* 29:1682–1684.
- 22 Kardos M, Åkesson M, Fountain T, Flagstad Ø, Liberg O, Olason P, Sand H, Wabakken P, Wikenros C,
- Ellegren H. 2018. Genomic consequences of intensive inbreeding in an isolated wolf population. *Nat Ecol Evol* 2:124–131.
- Karsten M, van Vuuren BJ, Goodman P, Barnaud A. 2011. The history and management of black rhino in
 KwaZulu-Natal: a population genetic approach to assess the past and guide the future. *Anim. Conserv.* 14:363–370.
- 28 Kenya Wildlife Service. 2021. National Wildlife Census 2021 Report. Available from:
- 29 https://kws.go.ke/content/national-wildlife-census-2021-report.
- Korneliussen TS, Albrechtsen A, Nielsen R. 2014. ANGSD: Analysis of Next Generation Sequencing Data.
 BMC Bioinformatics 15:356.
- Kotzé A, Dalton DL, du Toit R, Anderson N, Moodley Y. 2014. Genetic structure of the black rhinoceros
 (Diceros bicornis) in south-eastern Africa. *Conserv. Genet.* 15:1479–1489.
- Li H, Durbin R. 2009. Fast and accurate short read alignment with Burrows-Wheeler transform. *Bioinformatics* 25:1754–1760.
- Liu H, Prugnolle F, Manica A, Balloux F. 2006. A geographically explicit genetic model of worldwide human settlement history. *Am. J. Hum. Genet.* 79:230–237.
- Mak SST, Gopalakrishnan S, Carøe C, Geng C, Liu S, Sinding M-HS, Kuderna LFK, Zhang W, Fu S, Vieira
 FG, et al. 2017. Comparative performance of the BGISEQ-500 vs Illumina HiSeq2500 sequencing platforms for
 palaeogenomic sequencing. *Gigascience* 6:1–13.
- Manica A, Prugnolle F, Balloux F. 2005. Geography is a better determinant of human genetic differentiation
 than ethnicity. *Hum. Genet.* 118:366–371.

- Mantel N. 1967. The detection of disease clustering and a generalized regression approach. *Cancer Res.* 27:209–220.
- Meisner J, Albrechtsen A. 2018. Inferring Population Structure and Admixture Proportions in Low-Depth NGS
 Data. *Genetics* 210:719–731.
- 5 Minh BQ, Schmidt HA, Chernomor O, Schrempf D, Woodhams MD, von Haeseler A, Lanfear R. 2020. IQ-
- TREE 2: New Models and Efficient Methods for Phylogenetic Inference in the Genomic Era. *Mol. Biol. Evol.*37:1530–1534.
- 8 Moodley Y, Russo I-RM, Dalton DL, Kotzé A, Muya S, Haubensak P, Bálint B, Munimanda GK, Deimel C,
- 9 Setzer A, et al. 2017. Extinctions, genetic erosion and conservation options for the black rhinoceros (*Diceros*
- 10 *bicornis*). *Sci. Rep.* 7:41417.
- 11 Moodley, Yoshan, Isa-Rita M. Russo, Jan Robovský, Desiré L. Dalton, Antoinette Kotzé, Stéve Smith, Jan
- 12 Stejskal, et al. 2018. "Contrasting Evolutionary History, Anthropogenic Declines and Genetic Contact in the
- 13 Northern and Southern White Rhinoceros (*Ceratotherium Simum*)." Proceedings. Biological Sciences / The
- 14 *Royal Society* 285 (1890). https://doi.org/10.1098/rspb.2018.1567.
- 15 Moodley Y, Westbury MV, Russo I-RM, Gopalakrishnan S, Rakotoarivelo A, Olsen R-A, Prost S, Tunstall T,
- 16 Ryder OA, Dalén L, et al. 2020. Interspecific Gene Flow and the Evolution of Specialization in Black and White
- 17 Rhinoceros. Mol. Biol. Evol. 37:3105–3117.
- Moore AE, Larkin PA. 2001. Drainage evolution in south-central Africa since the breakup of Gondwana. *South Afr. J. Geol.* 104:47–68.
- Moore AE, (woody) Cotterill FP, Eckardt FD. 2012. The evolution and ages of Makgadikgadi palaeo-lakes:
 Consilient evidence from Kalahari drainage evolution South-Central Africa. *South Afr. J. Geol.* 115:385–413.
- 22 Muya SM, Bruford MW, W.-T. Muigai A, Osiemo ZB, Mwachiro E, Okita-Ouma B, Goossens B. 2011.
- 23 Substantial molecular variation and low genetic structure in Kenya's black rhinoceros: implications for
- conservation. Conserv. Genet. 12:1575–1588.
- Pebesma, E.J., R.S. Bivand. 2005. Classes and methods for spatial data in R. Available from: https://cran.r project.org/doc/Rnews/.
- Petkova D, Novembre J, Stephens M. 2016. Visualizing spatial population structure with estimated effective
 migration surfaces. *Nat. Genet.* 48:94–100.
- 29 Pickford M, Senut B, Hadoto D. 1993. Geology and palaeobiology of the Albertine Rift valley, Uganda-Zaire.
- **30** Volume I : geology. *Publication occasionnelle Centre international pour la formation et les échanges*
- 31 *géologiques* [Internet]. Available from: https://pascal-
- 32 francis.inist.fr/vibad/index.php?action=getRecordDetail&idt=6369901
- Quinlan AR, Hall IM. 2010. BEDTools: a flexible suite of utilities for comparing genomic features.
 Bioinformatics 26:841–842.
- 35 R Core Team. 2022. R: A language and environment for statistical computing. R Foundation for Statistical
- 36 Computing, Vienna, Austria Available from: https://www.r-project.org/
- Richard A. Becker OSC, by Ray Brownrigg. ARWRV. 2018. mapdata: Extra Map Databases. Available from:
 https://CRAN.R-project.org/package=mapdata
- Richard A. Becker OSC, by Thomas P Minka ARWRV by RBE, Deckmyn. A. 2018. maps: Draw Geographical
- 40 Maps. Available from: https://CRAN.R-project.org/package=maps
- 41 Riedel F, Henderson ACG, Heußner K-U, Kaufmann G, Kossler A, Leipe C, Shemang E, Taft L. 2014.
- 42 Dynamics of a Kalahari long-lived mega-lake system: hydromorphological and limnological changes in the
- 43 Makgadikgadi Basin (Botswana) during the terminal 50 ka. *Hydrobiologia* 739:25–53.

- 1 Roger S. Bivand, Edzer Pebesma, Virgilio Gomez-Rubio. 2013. Applied spatial data analysis with R, Second
- 2 edition. Springer, NY.
- Rookmaaker K. 2011. A review of black rhino systematics proposed in Ungulate Taxonomy by Groves and
 Grubb (2011) and its implications for rhino conservation. *Pachyderm* 50:72–76.
- Rookmaaker K, Antoine P-O. 2012. New maps representing the historical and recent distribution of the African
 species of rhinoceros: *Diceros bicornis*, *Ceratotherium simum* and *Ceratotherium cottoni*. *Pachyderm* 52:91–96.
- 7 Sánchez-Barreiro F, Gopalakrishnan S, Ramos-Madrigal J, Westbury MV, de Manuel M, Margaryan A, Ciucani
- 8 MM, Vieira FG, Patramanis Y, Kalthoff DC, et al. 2021. Historical population declines prompted significant
- 9 genomic erosion in the northern and southern white rhinoceros (Ceratotherium simum). *Mol. Ecol.* 30:6355–
- **10** 6369.
- Scharning K. Google Maps API v3 Tool. www.birdtheme.org/useful/v3tool.html [Internet]. Available from:
 http://www.birdtheme.org/useful/v3tool.html
- 13 Schubert M, Ermini L, Der Sarkissian C, Jónsson H, Ginolhac A, Schaefer R, Martin MD, Fernández R, Kircher
- 14 M, McCue M, et al. 2014. Characterization of ancient and modern genomes by SNP detection and phylogenomic
- and metagenomic analysis using PALEOMIX. *Nat. Protoc.* 9:1056–1082.
- Schubert M, Lindgreen S, Orlando L. 2016. AdapterRemoval v2: rapid adapter trimming, identification, and
 read merging. *BMC Res. Notes* 9:88.
- 18 Shafer ABA, Wolf JBW, Alves PC, Bergström L, Bruford MW, Brännström I, Colling G, Dalén L, De Meester
- L, Ekblom R, et al. 2015. Genomics and the challenging translation into conservation practice. *Trends Ecol. Evol.* 30:78–87.
- Skotte L, Korneliussen TS, Albrechtsen A. 2013. Estimating individual admixture proportions from next
 generation sequencing data. *Genetics* 195:693–702.
- 23 Theissinger K, Fernandes C, Formenti G, Bista I, Berg PR, Bleidorn C, Bombarely A, Crottini A, Gallo GR,
- Godoy JA, et al. 2023. How genomics can help biodiversity conservation. *Trends Genet*. [Internet]. Available
- 25 from: http://dx.doi.org/10.1016/j.tig.2023.01.005
- Van Coeverden de Groot PJ, Putnam AS, Erb P, Scott C, Melnick D, O'Ryan C, Boag PT. 2011. Conservation
 genetics of the black rhinoceros, Diceros bicornis bicornis, in Namibia. *Conserv. Genet.* 12:783–792.
- Veldkamp A, Schoorl JM, Wijbrans JR, Claessens L. 2012. Mount Kenya volcanic activity and the Late
 Cenozoic landscape reorganisation in the upper Tana fluvial system. *Geomorphology* 145-146:19–31.
- Waples RK, Albrechtsen A, Moltke I. 2019. Allele frequency-free inference of close familial relationships from
 genotypes or low-depth sequencing data. *Mol. Ecol.* 28:35–48.
- 32 Westbury MV, Hartmann S, Barlow A, Wiesel I, Leo V, Welch R, Parker DM, Sicks F, Ludwig A, Dalén L, et
- al. 2018. Extended and Continuous Decline in Effective Population Size Results in Low Genomic Diversity in
 the World's Rarest Hyena Species, the Brown Hyena. *Mol. Biol. Evol.* 35:1225–1237.
- 35 Westbury MV, Thompson KF, Louis M, Cabrera AA, Skovrind M, Castruita JAS, Constantine R, Stevens JR,
- 36 Lorenzen ED. 2021. Ocean-wide genomic variation in Gray's beaked whales, Mesoplodon grayi. *R. Soc. Open*
- 37 *Sci.* [Internet] 8. Available from: https://royalsocietypublishing.org/doi/10.1098/rsos.201788
- 38 Wickham H. 2016. ggplot2: Elegant Graphics for Data Analysis. Available from: https://ggplot2.tidyverse.org
- Yang Z. 2007. PAML 4: phylogenetic analysis by maximum likelihood. *Mol. Biol. Evol.* 24:1586–1591.
- 40 Zukowsky, L. Die Systematik der Gattung *Diceros* Gray, 1821. 1965. *Zool. Gart.* 30, 1–178.
- 41

1 Figure captions

2

Figure 1. The wide historic range of the black rhinoceros in sub-Saharan Africa and sampling locations. The shaded area indicates the historic range of distribution of the black rhinoceros (from Rookmaaker and Antoine 2012). Coloured dots represent the sampling locations of 80 georeferenced samples in our dataset, 53 historic and 27 modern. An additional eight samples lacked coordinates, but their country of origin was known. Two samples were of unknown origin (Table 1 and Table S1). Dot sizes represent the number of samples collected at each location.

9

10 **Figure 2.** Range-wide population genomic structure of black rhinoceros historic and modern sample sets. A) PCA of all 63 historic genomes coloured by country of origin. B) PCA of historic genomes sampled north of the 11 Zambezi River. C) PCA of historic and modem (all) genomes sampled north of the Zambezi River. D) PCA of 12 historic genomes sampled in southern Africa, south of the Zambezi River. E) PCA of historic and modern (all) 13 genomes sampled in southern Africa, south of the Zambezi River. F) Admixture analysis of historic individuals 14 showing range-wide population structure at K = 6. Values of $K \ge 10$, including modern genomes, are available 15 in Figures S4 and S6. G) Fossil calibrated phylogenomic tree using a single individual representative per 16 population. Branch labels show bootstrap values, gene concordance factors, and site concordance factors 17 18 respectively.

19

Figure 3. *Isolation by distance across the historic range of the black rhinoceros.* (A) Mantel regression showing the significant relationship between pairwise genetic and geographic distances for all georeferenced historic black rhinoceros. (B) Distribution of D-statistic values showing the relative genetic distance between the central and eastern populations (CE or EA) to either the southern (S) or northwestern population (NW). A negative D-score (blue) indicates a closer relationship to S, whereas a positive D-score indicates a closer relationship to NW (yellow).

26

Figure 4. Effective migration across the historic range of the black rhinoceros and summary of the inferred historic population structure in the black rhinoceros. The effective migration surface was inferred with EEMS (Petkova et al. 2016) based on genome-wide data from 53 georeferenced historic samples. The colour gradient represents effective migration rates in logarithmic scale; blue shades indicate rates higher than average, while grey shades represent migration rates lower than average. The six inferred historic populations were mapped onto the migration surface to determine their geographical distribution. Roman numerals denote regions of low (I-IV) and high (V-X) migration described in the text. Dot size represents the number of samples from each location.

34

Figure 5. Individual genomic diversity across geographically informed populations of black rhinoceros. A)
Individual genome-wide heterozygosity (GWhet) for 83 modern and historic samples, B) GWhet based on
geographical distribution for 53 georeferenced historic samples, C) Distribution of individual F_{RoH} values with a
window size of 1Mb and larger for 83 modern and historic samples per group is visualised. D) The geographic

1 distribution of F_{RoH} with a window size of 1Mb and larger is shown for 53 historic, georeferenced samples.

Figure 6. *Inbreeding through time and space*. Violin plots of individual percentages of genome in RoH across the
six major black rhinoceros populations were divided into size classes to investigate inbreeding at three sequential
timeframes of the recent past. Allowing for a generation time of 24 years equates to inbreeding between 517 –
1032 years (A), 207 - 516 years (B) and 0 - 206 years (C) for the small, medium and large F_{RoH} size classes
respectively. CE, Central Africa; EA, East Africa; NE, Northeastern; NW, Northwestern, RU, Ruvuma; S,
Southern.

9













